Asian Fisheries Science 11(1998):231-238 Asian Fisheries Society, Manila, Philippines https://doi.org/10.33997/j.afs.1998.11.3-4.006

# Effects of Hydrogen Sulfide on the Early Developmental Stages of Javanese Carp, *Puntius gonionotus* (Bleeker)

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#### Abstract

The toxicity of hydrogen sulfide (H<sub>2</sub>S) to the larvae and eggs of the Javanese carp, *Puntius gonionotus* (Cyprinidae), was determined using a flow-through bioassay. The eggs and larvae showed similar low tolerances to H<sub>2</sub>S. The EC<sub>50</sub> value for the eggs was 2.8 µg l<sup>-1</sup> H<sub>2</sub>S and the 96h-LC<sub>50</sub> (median lethal concentration) for the larvae was 2.7 µg l<sup>-1</sup> H<sub>2</sub>S. One hundred percent of the eggs failed to hatch at H<sub>2</sub>S levels higher than 3.4 µg l<sup>-1</sup> H<sub>2</sub>S, while a 100% mortality of the larvae occurred at 5.4 µg l<sup>-1</sup> H<sub>2</sub>S. The growth rate of the larvae exposed to H<sub>2</sub>S level higher than 0.6 µg l<sup>-1</sup> was significantly reduced. Newly hatched larvae from eggs exposed to more than 0.6 µg l<sup>-1</sup> H<sub>2</sub>S showed tail deformities. Histological studies showed increasing degrees of hyperplasia in fish gills at increasing H<sub>2</sub>S concentra-tions. This study indicates that *P. gonionotus* is vulnerable to organic pollution in the aquatic environment.

## Introduction

Hydrogen sulfide (H<sub>o</sub>S) is a toxic compound that occurs in natural waters as a result of the anaerobic decomposition of organic matter in sediments, domestic sewerage, pulp-mill wastes and other industrial effluents. In thermally stratified lakes, high concentrations of  $H_2S$  are found in hypolimnetic waters (Reynolds and Haines 1980). Many species of fish spend all or part of their life in shallow waters and may experience sulfide toxicity near the bottom. Many fishes lay eggs and spend their early and critical development stages at the bottom. Smith and Oseid (1974) reported that H<sub>2</sub>S levels commonly occurring in nature could be lethal to eggs and larvae of several freshwater fish species. Lethal concentrations of H<sub>o</sub>S would limit fish survival and reduce fishery production (Smith et al. 1976). Porter et al. (1986) reported that sulfide level in the aged pond sediment could be as high as 500  $\mu$ g l<sup>-1</sup>. Unionized hydrogen sulfide above  $0.3 - 5.0 \ \mu g \ l^{-1} H_{o}S$  is toxic to freshwater catfish Mystus

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nemurus, Bagridae (Hoque et al., unpubl. data) and shrimp Penaeus monodon, Penaeidae (Law 1988).

Little work has been done on  $H_2S$  toxicity in tropical fish (Law 1988; Hoque *et al.* 1996a, 1996b; Hoque *et al.* 1998) compared to fishes in temperate region (Bonn and Follis 1967; Aldelman and Smith 1970, 1972; Smith and Oseid 1972, 1975; Broderius *et al.* 1977; Reynolds and Haines 1980; Bagarinao and Vetter 1993). As  $H_2S$  is toxic and frequently occurs in fish ponds, this study was carried out to determine the acute and chronic toxicity of  $H_2S$  to *Puntius gonionotus* (Bleeker), Cyprinidae, an important commercial species in Malaysia, in its early developmental stages.

## **Materials and Methods**

## Test Organisms

*P. gonionotus* larvae were obtained from the Bukit Tinggi Fish Breeding Station (Department of Fisheries) in Pahang, Malaysia. The larvae were acclimatized for one week in the Aquatic Ecology Laboratory in Universiti Putra Malaysia (UPM) before they were used in the experiments. The mean weight and total length of the larvae were  $0.05\pm0.02$  g and  $1.6\pm0.2$  cm, respectively. The one-hour-old fertilized eggs were collected from the Fish Hatchery Unit, UPM, and used immediately for the toxicity tests.

#### Experimental Set-up and Hydrogen Sulfide Assay

A flow-through bioassay system for hydrogen sulfide was employed in this study (Law and Yusoff, unpubl. method). Under this system, hydrogen sulfide levels were maintained within 10% of the mean value throughout the experimental period. Hydrogen sulfide stock solutions in the ranges between 1,000-10,000 mgl<sup>-1</sup> total hydrogen sulfide, in oxygen-free deionized water at pH 9 were prepared daily and kept at 4°C. The analytical grade sodium sulfide  $(Na_2S.9H_2O)$  (Fluka, Switzerland) was used as the source of H<sub>2</sub>S. The stock solutions were pumped into bioassay flasks (1-l Erlenmeyer flasks) using DESAGA peristaltic pumps and instantly mixed with a large volume of seasoned oxygenated water at room temperature. To avoid rapid disappearance of hydrogen sulfide from the bioassay flasks, no aeration was given to the test solutions. The incoming oxygenated seasoned water was able to maintain the dissolved concentrations of the test solutions higher than 5 mgl<sup>-1</sup>. The total hydrogen sulfide in the test solutions was analyzed daily according to the standard method and the unionized hydrogen sulfide concentrations were calculated from pH and the equilibrium constant (Cline 1969; APHA-AWWA-WPCF 1989; Boyd 1990). Water quality parameters such as temperature, pH and dissolved oxygen were monitored routinely throughout the experiments to ensure that water quality was acceptable for the growth of the fish larvae. Dissolved oxygen was measured with a YSI oxygen meter (Model 57) and pH with a Corning pH meter (Model SA520). No significant difference (p>0.05) of water temperature (means ranged from 27.8-28.0 °C), pH (7.13-7.57) and dissolved oxygen (7.0-8.1

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mgl<sup>1</sup>) values in different treatments were detected throughout the study period. During the experiment, sulfide concentrations fluctuated at  $\pm 18.9$ . According to Sprague (1990), fluctuations of toxicant concentration within  $\pm 20\%$  of the mean during flow-through bioassays are acceptable.

## Acute Toxicity Test on the Eggs

The toxicity tests on eggs were done in 500-ml beakers using the flowthrough bioassay. Thirty one-hour fertilized eggs were randomly spread on the bottom of each beaker. The H<sub>2</sub>S concentrations in this test were 0.11±0.05, 0.62±0.16, 2.03±0.52 and 3.39±0.74 µg l<sup>-1</sup> H<sub>2</sub>S. A treatment without hydrogen sulfide served as a control. All treatments were in triplicates. The hatching rate in each beaker was recorded. The experiment was conducted for four days. The EC<sub>50</sub> value to determine the hydrogen sulfide concentration at which 50% of the eggs failed to hatch was calculated using the probit method (APHA-AWWA-WPCF 1989).

#### Acute and Chronic Tests on the Larvae

The acute and chronic tests on the fish larvae were conducted in 1-l Erlenmeyer flasks, using the flow-through bioassay system. The H<sub>2</sub>S concentrations were 2.07±0.61, 2.77±0.22, 3.80±0.67 and 5.35±0.72 µg l<sup>-1</sup> H<sub>2</sub>S in the acute toxicity tests and 0.03±0.01, 0.61±0.12, 1.90±0.35 and 3.41±0.50 µg l<sup>-1</sup> H<sub>2</sub>S in the chronic tests. The controls did not contain any hydrogen sulfide. Ten larvae were stocked in each flask at a total weight not exceeding one gram per liter of water. All the five treatments with three replicates were run at the same time using DESAGA peristaltic pumps to maintain the hydrogen sulfide concentrations. The acute toxicity experiment of the larvae ran for four days while the chronic toxicity test ran for 22 d.

No feeding was given to the larvae during the acute toxicity tests. In the chronic toxicity tests, the larvae were fed *ad libitum* daily with micropellets. Fish mortality was recorded at six-hour intervals and dead fish were removed immediately. Fish were considered dead when they showed no respiratory movements and no response to gentle prodding. The  $LC_{50}$  values were determined using the probit method. Both total length and weight of the fish larvae in the chronic toxicity tests were measured to determine their growth rates. Total length of the fish was measured to the nearest 0.1 cm and the body weight to the nearest 0.01 g. The differences in growth rates in the various treatments were analyzed using ANOVA and Duncan Multiple Range test according to Zar (1984) using Statistical Analysis System (SAS) computer software.

#### Histopathological Examination

Fish gills were dissected after 22 days of exposure to different sublethal doses of  $H_2S$ , or no  $H_2S$ . Gills were immediately preserved in 10% buffered formalin. They were then decalcified and processed by standard histopathological techniques, sectioned at 4  $\mu$ m, then stained with hematoxylin and eosin.

## **Results and Discussion**

The acute toxicity tests revealed that  $H_2S$  was highly toxic to eggs and larvae of *P. gonionotus*. One hundred percent mortality of eggs and larvae was found at  $H_2S$  levels higher than 3.4 µg  $l^{-1}$  and 5.4 µg  $l^{-1}$ , respectively (Figs. 1 and 2). The EC<sub>50</sub> value of  $H_2S$  for the eggs was 2.8 µg  $l^{-1}$  while the 96-h LC<sub>50</sub> for larvae was 2.7 µg  $l^{-1}$   $H_2S$ . The effects of  $H_2S$  is generally more severe on fish larvae than on fish eggs. This is true for goldfish (*Carassius auratus*, Cyprinidae) and fathead minnows (*Pimephelas promelas*, Cyprinidae) (Smith and Oseid 1974), and northern pike (*Esox lucius*, Esocidae) (Bonn and Follis 1967). However, a reverse phenomenon was observed for goldfish and bluegills (*Lepomis macrochirus*, Centrarchidae), where the EC<sub>50</sub> and 96-h LC<sub>50</sub> values of  $H_2S$  for eggs and larvae were 20 µg  $l^{-1}$  and 25 µg  $l^{-1}$ , and 16 µg  $l^{-1}$  and 30 µg  $l^{-1}$ , respectively (Smith and Oseid 1974). In this study, eggs of larvae of *P*.



Fig. 2. Percentage of larval mortality after 96-hour exposure to different concentrations of  $H_{y}S$ .

gonionotus were equally sensitive to  $H_2S$ . The results also indicated that *P. gonionotus* was more sensitive to hydrogen sulfide than other fishes. The EC<sub>50</sub> and LC<sub>50</sub> of  $H_2S$  for temperate fish species are considerably higher, probably due to the lower temperatures (9-21°C) used in the former than in the present study (28°C). Adelman and Smith (1972) reported that





96-hr median tolerance limits (TL<sub>50</sub>) value for  $H_2S$  increased from 4 µg l<sup>-1</sup>  $H_2S$  at 25°C to 530 µg l<sup>-1</sup>  $H_2S$  at 6.0°C in the goldfish. The 96-h  $LC_{50}$  for the fathead minnow and brown trout (*Salmo trutta*, Salmonidae) larvae was 7.0 µg l<sup>-1</sup>  $H_2S$  at 24°C and 13°C, respectively (Smith and Oseid 1975; Reynolds and Haines 1980).

In the chronic flow-through study with sublethal  $H_2S$  concentrations, there was a significant reduction (p<0.05) in fish growth rates when the  $H_2S$ level was higher than 0.03 µg l<sup>-1</sup>  $H_2S$  (1% of  $LC_{50}$ ) (Fig. 3). At the highest concentration of 3.41 µg l<sup>-1</sup>  $H_2S$ , fish weights increased by only about 20% after 22 d while the control fish grew by 93%. This is in agreement with Smith and



Fig. 3. Percentage increase in body weight of larvae exposed for 22 days to different concentrations of  $H_2S$ .

Oseid (1976), who reported that growth retardation in juvenile bluegills occurred even at 2.4  $\mu$ g l<sup>-1</sup> H<sub>2</sub>S.

The no-effect level of H.,S growth Ρ. the of on gonionotus in this study was about 0.03 µg l<sup>-1</sup> H<sub>o</sub>S. Smith and Oseid (1975) have examined the effects of chronic toxic levels of H<sub>o</sub>S on the early life stages of various freshwater fish species and found that a safe level of H<sub>2</sub>S which would ensure survival and growth of fish populations would be between 2 to 4  $\mu$ g l<sup>-1</sup> H<sub>2</sub>S at 20°C. Since the no-effect level of  $H_{2}S$  to P. gonionotus is

much lower than this value, any form of organic pollution in the aquatic environment would have adverse impacts on the fish population.

More than 30% of the eggs did not hatch when exposed to more than 0.6  $\mu$ g l<sup>-1</sup> H2S. Those that did hatch had tail deformities (Fig. 4). Similar vertebral malformations have been observed in the sheephead minnow (*Cyprinodon variegatus*, Cyprinodontidae) after a 28-day exposure to the herbicide triflura-

lin (Couch *et al.* 1979) and in other fish, due to chemical contaminants in general (Moore and Hixson 1977; Muramoto 1981).

Fig. 4. Deformed tail of newlyhatched larva from *Puntius* gonionotus egg exposed for 96 h to hydrogen sulfide higher than 0.6 µg l<sup>-1</sup> H<sub>2</sub>S.



Gills of *P. gonionotus* exposed to  $H_2S$  exhibited various degrees of epithelial hyperplasia, even at the lowest concentration (0.03 µg l<sup>-1</sup> H<sub>2</sub>S). The gill lamellae of the unexposed larvae are shown in Fig. 5. Hyperplasia was generally more pronounced at the base of the secondary lamellae and the degree of hyperplastic changes increased with  $H_2S$  concentrations (Fig. 6). Slight epithelial lifting was also observed in fish exposed to  $H_2S$ . The gill lamellae of the exposed fish were congested with blood cells, indicating irritation from  $H_2S$  exposure. Epithelial hyperplasia and separation of the epithelial layer from supportive tissues usually mean disorders in gill function that affect the physiology or cause the death of fish (Eller 1975; Gardner 1975; Smart 1976). In addition, hyperplasia and separation of the epithelium are associated with asphyxiation, partial or complete loss of secretory or excretory function, impairment of oxygen-carbon dioxide exchange, and loss of plasma electrolytes or proteins (Smith and Piper 1972; Mitchell *et al.* 1978).

This study illustrates that  $H_2S$  is toxic in the early developmental stages of *P. gonionotus*, even at low concentrations of less than 2 µg l<sup>-1</sup> H<sub>2</sub>S. Smith *et al.* (1976) suggested that any detectable concentrations of hydrogen sulfide are detrimental to fish production. Since sulfide is commonly found in ponds with high accumulations of organic matter, fish farm managers must include ways to prevent or overcome anaerobic conditions in the bottom of ponds to avoid sulfide build-up to toxic levels.



Fig. 5. Normal gill lamellae of P. gonionotus not exposed to  $H_2S$ .

Fig. 6. Hyperplasia at the base and along the secondary gill lamellae of *P. gonionotus* larvae exposed to 3.41µg l<sup>+</sup>H<sub>s</sub>S.



This study was supported by the Malaysian Government Research Grant (IRPA-Intensification of Research Priority Areas) project no. 1-07-05-078.

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